

# Indoor and outdoor exposure to PM<sub>2.5</sub> during COVID-19 lockdown in suburban Malaysia

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During the COVID-19 pandemic, key policies aimed at reducing exposure to the virus from social distancing, restrictions on travel through to strongly enforced lockdowns. However, COVID-19 restrictions required people to spend more time at home so the exposure to air pollutants shifted to being derived from that of domestic interiors, rather than outdoors or the workplace environment. This study aims to characterise the influence of lockdown intervention on the balance of indoor and outdoor PM<sub>2.5</sub> exposure in a Malaysian suburb. We also calculate the potential health risk from exposure to both indoor and outdoor PM<sub>2.5</sub> to give context to personal exposure assessment in different microenvironments during the COVID-19 lockdown, known locally as *Movement Control Orders (MCO)*. The implementation of the MCOs slightly reduced daily average of outdoor PM<sub>2.5</sub> concentrations (median of 12.63  $\mu\text{g m}^{-3}$  before and 11.72  $\mu\text{g m}^{-3}$ ). In the Malaysian apartment considered here, cooking led to a substantial increase in exposure from increasing concentrations in PM<sub>2.5</sub> during a COVID-19 lockdown (maximum average concentration at 52.2  $\mu\text{g m}^{-3}$ ). The estimated excess risk to health was about 25% for lung cancer from staying indoor. Thus, there seems a potential for greater exposure to fine particles indoors under lockdown, so it is likely premature to suggest that more lives were saved through a reduction of outdoor pollutants than lost in the pandemic. Unfortunately, little is known about the toxicity of indoor particles and the types of exposures that result where people increase the amount of time they spend working from home or staying indoors, especially during periods of lockdown.

**Keywords:** Cancer, Cardio-respiratory diseases; Cooking; Indoor air quality; Lockdown

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## 42 INTRODUCTION

43

44 During the COVID-19 pandemic lower concentrations of ambient air pollutants were observed  
45 at many areas of the world due to widespread restrictions on travel, social activities and work.  
46 People spent much of their lives indoors, which further shifted human exposure to pollutants during  
47 periods of restricted activity. The pandemic infections were caused by a zoonotic virus of the  
48 SARS-CoV and MERS-CoV families (Mackenzie and Smith, 2020) first reported in Wuhan, China  
49 (Zhu et al., 2020). The COVID-19 crisis affected some six and a half million people worldwide  
50 with an 11% mortality rate through the first six months of 2020 (WHO, 2020). In the effort to  
51 prevent further outbreak, governments around the globe implemented restrictive measures as part  
52 of the COVID-19 containment.

53 Imposed changes to human activity led to a reduction in pollutant emissions, which was  
54 unprecedented on such a wide geographic scale. This provided a unique opportunity for researchers  
55 to assess the effect of the changes on air quality as measured from data monitored from satellite to  
56 ground measurements (Dutheil et al., 2020; Wang and Su, 2020; Xu et al. 2020; Yue et al., 2020;  
57 Sharma et al., 2020; Abdullah et al., 2020; Tobías et al., 2020; Muhammad, et al., 2020; Dantas et  
58 al., 2020; Nakada and Urban, 2020; Chauhan and Singh, 2020). Prior to the COVID-19 pandemic,  
59 transportation and the industrial sector were large and often growing sources of air pollution. The  
60 reduction of these sources during lockdown frequently improved air quality, with a notable  
61 decrease in NO<sub>2</sub> concentrations, although there have been increases in ozone concentrations e.g.  
62 across Northern China surface O<sub>3</sub> concentrations rose during the epidemic (Huang et al., 2020; Shi  
63 and Brasseur, 2020), with large increases in urban Beijing and Wuhan and more modest changes  
64 in Shanghai and Guangzhou (Zhao et al., 2020).

65 The strict lockdown in China limited human mobility, suspending intra city transport and  
66 closed factories. The air quality index improved 7.8% (Bao and Zhang, 2020), with the PM<sub>2.5</sub>

67 concentration reduced by  $>30 \mu\text{g m}^{-3}$  in Wuhan (Wang et al., 2020), although  $\text{SO}_2$  and CO were  
68 not greatly reduced during the COVID-19 lockdown as goods transport, coal-fired power plants  
69 and domestic heating were still needed in China. Overall these reductions may have been more  
70 subtle than often proposed in the media (Brimblecombe and Lai, 2020a; Cole et al. 2020; Silver,  
71 2020), but they may have shifted the weekly and diurnal pattern of pollutants (Brimblecombe and  
72 Lai, 2020b). During lockdown in Malaysia there were decreases in concentrations of particulate  
73 matter in many locations, although there was evidence of illegal local biomass burning activities  
74 by individuals and private companies (Abdullah et al. 2020; Kanniah et al. 2020; Mohd Nadzir et  
75 al., 2020). Elsewhere increased outdoor burning activities of garden and household waste during  
76 lockdown in London came as people used the period at home to do spring cleaning (LFC, 2020).  
77 Increases in particulate matter may also come about through the production of secondary aerosol  
78 which arises from increased ozone and  $\text{NO}_3$  radical formation at night when  $\text{NO}_2$  concentrations  
79 were low (Huang et al., 2020a).

80 An initial study by Chen et al. (2020) found the reduction of  $\text{PM}_{2.5}$  during this period can lead  
81 to a 73% reduction in mortality risk from  $\text{PM}_{2.5}$  related deaths, but others said the pollution was  
82 not avoided (Wang et al., 2020). It is even suggested that health benefits related to  
83 cardiopulmonary disease can outweigh COVID-19 mortality under lockdown. Another health co-  
84 benefit from the 20% drop of air pollution levels is reduced asthma cases and risk of premature  
85 death (Venter et al., 2020). However, personal exposure to air pollution was quite different under  
86 the restrictions imposed to control COVID-19, as this often-involved urban populations spending  
87 almost all of their time indoors (Abouleish, 2020). Exposure in such microenvironments during  
88 COVID-19 have gone largely unreported, despite the enormous changes it imposed on daily lives.  
89 The preoccupation with outdoor air pollution (Huang et al., 2020b) may have distorted our views  
90 of how exposure might have changed. This is an especially distinct period when people spent so

91 little time outdoors. A proper assessment of personal exposure during lockdown should account for  
92 the heterogeneity of relevant microenvironments, the number of occupants, housing conditions,  
93 activities and lifestyle in the indoor settings. All these are likely to influence both exposure to  
94 indoor and outdoor pollutants that resulted from stay-at-home policies mandated under COVID-  
95 19. Indoor outdoor ratios of PM<sub>2.5</sub> are often greater than unity in residential settings (Cao et al.,  
96 2005; Huang et al., 2007) and Thakur et al. (2020) have raised concern over increased rate of  
97 cooking and smoking activities, with particles in the kitchen potentially rather toxic with respect to  
98 their oxidative capacity (Shao et al., 2007).

99 Health may be of special consent among vulnerable groups such as pregnant women, children,  
100 the elderly and people with underlying respiratory disease and immunodeficiency. In crowded  
101 interiors where the range of daily activity is much restricted there was likely more cooking,  
102 household repairs and hobbies, so indoor air quality was likely more relevant to health during  
103 COVID-19 restrictions than normal. Additionally, there are other, perhaps more serious health  
104 problems under lockdown: lack of access to medical support, mental illness from isolation  
105 (Venkatesh and Edirappuli, 2020), increased alcohol consumption (Clay and Parker, 2020),  
106 domestic violence (Malathesh et al., 2020) etc.

107 This study reports PM<sub>2.5</sub> concentrations measured outdoors and indoors at an apartment  
108 building in suburban Malaysia, with those retrieved from a nearby campus mini monitoring station  
109 and an official fixed monitoring site. After more than 1000 confirmed positive cases in Malaysia,  
110 Movement Control Orders (MCOs) were introduced in mid-March. Such lockdowns offer an  
111 appealing opportunity for experimental studies of air pollution, and can provide a causal  
112 understanding relevant to improved air quality in line with studies of pollution reduction during the  
113 2008 Beijing Olympics (Wang et al., 2010; He et al., 2016), street protests in Hong Kong  
114 (Brimblecombe and Ning, 2015; Brimblecombe, 2020) and driving restrictions in Tianjin and

115 Beijing (Zhang et al., 2020). This study aims to characterise the influence of MCO intervention on  
116 the balance of indoor and outdoor PM<sub>2.5</sub> exposure and likely potential changes in the risk to health,  
117 through increased times spent indoors under the lockdown restrictions.

118

## 119 **METHODS**

120

121

### 122 *Study Period*

123 Movement control in Malaysia was initially imposed from 18<sup>th</sup> March 2020 under the Prevention  
124 and Control of Infectious Diseases Act 1988 and the Police Act 1967 (MKN, 2020). Table 1 shows  
125 the different phases of MCO introduced by the Malaysian government in the fight to reduce  
126 COVID-19 infections. During the first phases of the MCOs (MCO1-MCO3), all Malaysians were  
127 advised to stay at home, maintain social distancing and stay within 10 km of their residential area  
128 when obtaining groceries or medication. Control was relaxed following advice by the Ministry of  
129 Health on 4<sup>th</sup> of May 2020, after the number of daily cases and active cases of COVID-19 declined.  
130 We compare PM<sub>2.5</sub> concentrations retrieved before MCO imposition (n<MCO=38 days) and after  
131 MCO periods that ran from MCO1 through to MCO5 (nMCO1-5=51 days) measured at the outdoor  
132 campus monitoring equipment (UPM) and the site at Putrajaya. These are compared with the  
133 separate set of measurements made as part of the indoor-outdoor apartment monitoring between  
134 MCO3 and MCO5 from 09:00 to 19:00 (n=13 days).

### 135 *Study Area*

136 The sampling sites in this study were (i) a mini monitoring platform in a university campus, (ii)  
137 an official measurement site of the Malaysia Department of Environment and (iii) inside and  
138 outside at an apartment located in suburban areas of Selangor, a state in the central region of  
139 Peninsular Malaysia (Fig. 1). Selangor covers an area of 7957 km<sup>2</sup> and has a population of over 6.3  
140 million. It consists of residential neighbourhoods and is home to a number of universities. The

141 university campus of Universiti Putra Malaysia is ~28 km southeast from the centre of Kuala  
142 Lumpur (Fig.1). The main outdoor sources of PM<sub>2.5</sub> at the UPM site is traffic on campus roads and  
143 nearby residential areas (~500 m distant). The apartment is located 11 km southwest from the  
144 campus and approximately 2 km from the highway. Our campus observations were made between  
145 February (before MCO) through May 2020 (during MCO), which spans the inter-monsoons from  
146 April to May in Malaysia.

#### 147 ***Ambient Measurements***

148 Ambient particulate matter measurements from the campus are taken over 85 days; before (9  
149 February-17 March 2020) and during (18 March-7 May 2020) the implementation of partial  
150 lockdown MCOs. This mini monitoring station was installed and managed by Enviro ExcelTech  
151 Sdn Bhd. We retrieved PM<sub>2.5</sub> data from the Aeroqual AQY-1 (Aeroqual, Auckland NZ), which is  
152 equipped with particulate (PM<sub>2.5</sub>), gaseous (NO<sub>2</sub> and O<sub>3</sub>), external temperature and humidity  
153 sensors. This PM<sub>2.5</sub> light-scattering optical particle sensor with RH correction ranges between 0 to  
154 1000 µg m<sup>-3</sup> and uses wireless technology to communicate its readings. The AQY-1 was housed in  
155 a sampling enclosure, mounted on a poll at 2-metre height with power supply via solar panels, with  
156 back up electricity and provides minute-by-minute PM<sub>2.5</sub> measurements. The AQY-1 has shown  
157 very good agreement (Karagulian *et al.*, 2019) with a reference system ( $R > 0.85$ ;  $0.8 < \text{slope} <$   
158  $1.2$ ). The values for Putrajaya come from a government monitoring site, which is located  
159 approximately 14 km from the campus site. PM<sub>2.5</sub> measurements were made using the Thermo  
160 Scientific TEOM 1405-DF, under maintenance by a private company, Transwater Sdn Bhd, which  
161 have been granted a 15 year concession to operate the site.

#### 162 ***Apartment Measurements***

163 The concentrations of PM<sub>2.5</sub> at the apartment building were measured using two units of TSI  
164 DustTrak II (Model 8532, TSI Inc., Shoreview, MN) and logged at a 1-minute time resolution.  
165 These light-scattering laser photometers were calibrated and validated by the manufacturer. The  
166 monitors were placed at the kitchen (indoor) and near the window (outdoor) in a vacant bedroom.  
167 Indoor and outdoor measurements were made simultaneously on 12 days during MCO3 (22 April-  
168 28 April 2020), MCO4 (1 May-3 May 2020) to MCO5 (4 May-7 May 2020) from 09:00 to 19:00  
169 (total sampling time 84 h). The newly constructed apartment unit covers an area of 787 square feet  
170 (i.e. 73 m<sup>2</sup>) and was built four years ago. The kitchen is equipped with a natural gas stove, hood  
171 exhaust duct and located within an open floor plan near the living area. All windows were opened  
172 in both kitchen and bedroom areas to maintain smooth air supply from outdoor and simulate normal  
173 conditions for a house in Malaysia where most use natural ventilation, although during cooking a  
174 fan with 1310 m<sup>3</sup> h<sup>-1</sup> suction could reduce the pollutant concentrations, as this flow would mean an  
175 air change rate 5 h<sup>-1</sup> for the entire apartment (242 m<sup>3</sup>).

#### 176 ***Traffic Data***

177 Daily traffic flow data were obtained from the exit toll stations near the UPM campus. The toll  
178 data was captured by site-based traffic sensors installed at each toll plaza and operated by the  
179 highway toll concessionaire or build-operate-transfer operator company, PLUS Malaysia Berhad.

#### 180 ***Estimation of Health Outcomes***

181 The relative risk (*RR*) is derived from the concentration of PM<sub>2.5</sub> measured at the campus and  
182 Putrajaya monitoring site and from the indoor apartment measurements and then applied to make  
183 the epidemiology-based excess risk (*ER*) calculations as shown in Eq. 2 (Kumar *et al.* 2020). The  
184 annual WHO standard of 10 µg m<sup>-3</sup> is used for the baseline PM<sub>2.5</sub> concentrations.

$$185 \quad RR = \exp [\beta(C - C_0)], \quad C > C_0 \quad (1)$$

$$186 \quad ER = RR - 1 \quad (2)$$

187 where  $\beta$ , the exposure-response coefficient was adopted from a linear dose response relationship  
188 used to estimate the health burden from acute respiratory diseases (age < 5), cardio-respiratory  
189 diseases (age > 30) and lung cancer (age > 30) as given by Ostro (2004) and Kwan *et al.* (2017).  
190 Here  $C$  is the average concentration that is measured from the site and  $C_0$  is the WHO threshold  
191 concentrations for PM<sub>2.5</sub>.

192

## 193 **RESULTS AND DISCUSSION**

194

### 195 *Ambient PM<sub>2.5</sub> before and during MCO*

196 The daily change in ambient PM<sub>2.5</sub> concentrations are shown in Fig. 2a, which depicts the daily  
197 average PM<sub>2.5</sub> concentrations before MCO (9 February-17 March 2020) and during MCO (18  
198 March-7 May 2020). The implementation of the MCOs slightly reduced daily average PM<sub>2.5</sub>  
199 concentrations at UPM, which had a median of 12.63  $\mu\text{g m}^{-3}$  before and 11.72  $\mu\text{g m}^{-3}$  under the  
200 MCOs (significant in Mann-Whitney test:  $U=1237$ ;  $p_1\sim 0.013$ ), while for Putrajaya the differences  
201 were less distinctive 17.39  $\mu\text{g m}^{-3}$  before and 16.25  $\mu\text{g m}^{-3}$  during (Mann-Whitney test:  $U=1107$ ;  
202  $p_1\sim 0.13$ ). With the exception of the first day of the record, the average concentrations of PM<sub>2.5</sub> did  
203 not exceed the standard of daily 24-h PM<sub>2.5</sub> of *The New Malaysia Ambient Air Quality Standard* of  
204 35  $\mu\text{g m}^{-3}$ , and the more stringent limit of *WHO Air Quality Guideline* of 25  $\mu\text{g m}^{-3}$ . Similarly,  
205 Ash'aari *et al.* (2020) observed the reduction of PM<sub>2.5</sub> at the sub-urban areas and did not exceed  
206 the guidelines during the different phases of MCO lockdown. A study by Kanniah *et al.* (2020)  
207 also revealed that PM<sub>2.5</sub> was higher than the guidelines in 2019 compared with those in 2020  
208 between March and April. The reduction of transport, and closure of educational institutions,  
209 government, and private agencies is generally believed to have caused the lowered PM<sub>2.5</sub>



210 concentrations during movement control (Mohd Nadzir *et al.*, 2020), although during the early part  
211 of lockdown pollutant concentrations were not greatly reduced (Abdullah *et al.*, 2020).

212 Decreased pollutant concentrations gain support from recorded traffic flow at the UPM road exit  
213 displayed over the period 9 February to 7 May 2020 (Fig. 2b) as it suggests that after lockdown  
214 traffic flows decreased rapidly and at least in terms of use of major highways citizens were  
215 compliant, and rapidly adapted to the new regime, despite widely held views that many took a long  
216 time to follow the regulation (Lim, 2020; Yusof, 2020). Even as lockdown, ended the return to  
217 normal was rather slow and as late as the end of June traffic was some 10% lower than before the  
218 MCOs had been imposed. Again, the media drew attention to rapid increases in traffic flow after  
219 the MCOs ended after May 3rd (TheStar, 2020), although there was a rise, it was clearly to nothing  
220 like the level typical before COVID-19. The dramatic changes in traffic flow are not well mirrored  
221 by the changes in PM<sub>2.5</sub>. This reminds us that there are many sources of pollution apart from  
222 highway traffic, and of course changing weather undoubtedly affected the concentrations during  
223 the MCOs. For example, high relative humidity levels were recorded during the MCO that may  
224 related to the rain events that can reduce PM<sub>2.5</sub> (Ash'aari *et al.* 2020). Any meteorological  
225 parameters do not influence the PM<sub>2.5</sub> levels before MCO.

226 Superficially the end of lockdown (MCO4 and MCO5) looks to have the lowest concentrations  
227 of PM<sub>2.5</sub>, despite the gradual increase in traffic, although the very last days of MCO5 showed  
228 increases in concentrations, though still less than 20  $\mu\text{g m}^{-3}$ . An ANOVA test revealed little  
229 difference in concentration between the various MCO periods. There were slightly higher  
230 concentrations, though statistically non-significant, during the weekends during the MCOs. This is  
231 rather the reverse of the normal situations where weekends typically have lower PM<sub>2.5</sub>  
232 concentrations. Such an outcome would be expected from lower weekend activity, such as the  
233 traffic flow illustrated by Fig. 2b, where the pairs of weekend days show lower flows across the

234 entire period, and there is no particular increase in traffic flow at the UPM exit due to people  
235 undertaking weekend shopping. It is possible that shopping under the MCOs was very localized,  
236 but this in itself would have in effect reduced the total burden of emissions. It should be noted that  
237 a wide variety of sources other than traffic exhaust may also contribute to the reduction of PM<sub>2.5</sub>  
238 including local biomass-burning activities. Ash'aari *et al.* (2020) suggest that the decreased level  
239 of PM<sub>2.5</sub> began during MCO4 due to reduction of fire emissions monitored from MODIS-derived  
240 hotspots and fire locations in Malaysia.

241

#### 242 ***Indoor-outdoor Measurements of PM<sub>2.5</sub> in the Apartment***

243 The PM<sub>2.5</sub> concentrations measured in the kitchen (indoor) and window (outdoor) during  
244 MCO3-MCO5, between 22 April until 8 May (n =13 days), are shown as points in Fig. 3. The  
245 amount of PM<sub>2.5</sub> steadily increased during the afternoons and constantly higher in the kitchen  
246 compared to the vacant room. Accounting for a typical daily meal preparation, the average modern  
247 adult spends over 10% of his or her during the day on a daily basis in the kitchen which may include  
248 eating and cleaning up from meals (Marc' *et al.*, 2018). However, considering the lower frequency  
249 of cooking before the lockdown for one person living in the apartment, the addition of extra cooking  
250 time each day during lockdown (approximately 20% of the day during meal preparation on food  
251 preparation, cooking, and cleaning) may represent an important source of indoor particles, and  
252 multiple exposures if the whole family is home. The distribution of PM<sub>2.5</sub> was higher during lunch  
253 time and gradually increased during preparation of the evening meal. It is noted that most cooking  
254 styles involved pan-frying and stir-frying for evening meal preparation, therefore showing a  
255 distinctive high peak between 17:30 to 18:30. Elevated concentrations were especially noted  
256 during MCO where a maximum concentration was observed in the evening at 52.2 µg m<sup>-3</sup>. Other  
257 cooking activities using the kitchen during the MCO involved boiling egg and chicken and baking.

258 The usage of gas stove either propane or natural gas has been suggested to contribute to the airborne  
259 particles related to the cooking method. The process of coagulation, condensation and evaporation  
260 will take place during cooking and influence the temporal variability in emissions (Huboyo *et al.*,  
261 2011). The emissions from cooking activities would be higher when the majority of people stayed  
262 at home the whole day so they may have been exposed to increased indoor air pollutants during the  
263 lockdown. Wan *et al.* (2011) have established the extent to which the average indoor particles of  
264 PM<sub>2.5</sub> and ultrafine concentrations were found higher than the background level in the living room.  
265 The dispersion of particles from the kitchen to the living room indicates that the health impact is  
266 not limited to occupants in the kitchen. In their study, the particulate emissions were found to  
267 disperse rapidly through the apartment and the particle number concentration can remain elevated  
268 for up to 90 minutes and as much as an hour in adjacent spaces after cooking in the kitchen. The  
269 average indoor-outdoor (I/O) ratio also shows consistently higher than 1 in our study. This may  
270 give indication on the elevation of indoor particles compared to outdoor. Other sources of indoor  
271 particles during stay at home are household dust, smoke from candles and cigarettes (DEFRA,  
272 2020).

273

#### 274 ***Exposure to PM<sub>2.5</sub> and Risk***

275 The average excess risk (ER) is shown in Fig. 4 where, ER (%) of the health burden from acute  
276 respiratory diseases (age < 5), cardio-respiratory diseases (age > 30) and lung cancer (age > 30)  
277 estimated from average PM<sub>2.5</sub> concentrations measured at indoor apartment, UPM (campus) and  
278 Putrajaya monitoring station during the lockdown under the MCOs. Estimated health burden before  
279 the MCO lockdown from UPM and Putrajaya measurements are also shown. Greater reduction in  
280 ER during lockdown as compared before lockdown is observed at UPM (3.4%) and Putrajaya  
281 (2.5%) for adult lung cancer risk compared to acute respiratory infection among children. These

282 results demonstrate that a lower excess risk was obtained during lockdown MCOs for long-term  
283 positive health impact. Children below than 5 years old living within UPM and Putrajaya is more  
284 likely to present acute health effects than in a group that is exposed to outdoor concentration of 10  
285  $\mu\text{g}/\text{m}^3$ , but people were of course indoors over the period of lockdown. Recent estimates by Giani  
286 *et al.* (2020) found that 10, 000 of premature deaths from air pollution exposure were avoided in  
287 China and Europe due to the reduction of  $\text{PM}_{2.5}$  during lockdown interventions. Other study that  
288 investigated the health and economic impact of lockdown across a few cities in India also found  
289 health and economic co-benefit due to lockdown across five Indian cities with decrease 30 to 50%  
290 in ER and avoided 630 premature deaths that cost 0.69 billion USD (Kumar *et al.* 2020).

291 Excess risk to health is found highest for lung cancer estimation when staying indoor (25.8%)  
292 during lockdown. This suggested that an individual in a group living in the indoor environment that  
293 is exposed to the corresponding  $\text{PM}_{2.5}$  concentration will encounter a raised health risk for mortality  
294 from lung cancer. It is important to note that the excess risk for cardiorespiratory mortality is also  
295 higher for indoor  $\text{PM}_{2.5}$  exposure (17.5%) during period of lockdown. Nonetheless, we noted that  
296 we only estimated the indoor exposure during lockdown period, therefore a comprehensive  
297 intervention of short and long-term health impacts should be accounted in the future study. The  
298 impact of particles in the indoor environment deserves further investigation.

### 299 ***Policy relevance***

300 Much has been made of the declining concentrations of some primary pollutants observed under  
301 lockdown with thoughts that this may provide guidance for future policies. Previous experiences  
302 suggest that the public are aware that air pollution returns once consumption activities resume and  
303 short-term restrictions are relaxed (Brimblecombe and Zong, 2019). A number of potential future  
304 pathways have been suggested (Bergman, 2020), but it is not clear how the patterns of human life

305 will change when the pandemic is over. Suggestions that lockdown decreases in traffic might be  
306 replicated in a post COVID-19 world may be difficult to achieve, and as observed here decreases  
307 in traffic were not paralleled by equivalent decreases in PM<sub>2.5</sub> concentrations. It has been argued  
308 that the experience of lockdown will encourage people to increasingly work from home in the  
309 future. However, this should raise concerns about the enhanced potential for indoor exposures and  
310 concomitant health risks. While there is some knowledge of a range of indoor microenvironments,  
311 our knowledge is often limited to simple concentration measurements, so much less is known of  
312 the health risk imposed by a range of different types of indoor particulate material.

313

## 314 **CONCLUSIONS**

315

316 Our study characterised the influence of MCOs lockdown intervention on the balance of indoor  
317 and outdoor PM<sub>2.5</sub> exposure and likely potential changes in the risk to health. The reduction in  
318 human movement and changed work patterns led to reduced pollutant emissions widely observed  
319 as a reduction in air pollutant concentrations. The daily median outdoor concentrations of PM<sub>2.5</sub>  
320 were reduced during the lockdown of 12.63  $\mu\text{g m}^{-3}$  before and 11.72  $\mu\text{g m}^{-3}$  under the MCOs at  
321 UPM (campus site) (Mann-Whitney test:  $U=1237$ ;  $p1\sim0.013$ ). Meanwhile, the reduction for  
322 Putrajaya was less distinctive 17.39  $\mu\text{g m}^{-3}$  before and 16.25  $\mu\text{g m}^{-3}$  during (Mann-Whitney test:  
323  $U=1107$ ;  $p1\sim0.13$ ). Our study suggested that cooking activities led to a substantial increase in PM<sub>2.5</sub>  
324 exposure during COVID-19 lockdown (maximum average concentration at 52.2  $\mu\text{g m}^{-3}$ ). Our  
325 estimation from health risk calculation has highlighted the relevance of staying indoor during  
326 lockdown with regard to health of the population. The excess risk to health is found at 25.8% for  
327 lung cancer estimation when staying indoor. Traditionally it maybe that only a few of the family  
328 were at home during cooking, but under MCOs it would mean exposure for the whole family.  
329 Indoor pollutants are often found at higher concentrations than outdoors (I/O ratio > 1) in our

330 apartment observation, so it is hardly surprising if extended periods spent indoors under lockdown  
331 increases exposure to particulate material. The widespread belief that exposures to air pollutants  
332 during COVID-19 were lower or that they drove susceptibility to infections seem unconvincing if  
333 these depend on only outdoor observations of concentrations. The simple set of measurements  
334 presented here would suggest that exposures in crowded interiors occupied for long periods was  
335 likely higher in terms of particulate concentrations. Little research has been done on examining the  
336 way people spent their lives under lockdown and to explore social disparities perhaps crucial in the  
337 way crowded living environments were utilised. Also, exposure would likely have been higher than  
338 outdoors, though it is far from certain whether the types of particles indoors under lockdown  
339 represented an enhanced risk to short term health. The toxicology of indoor particles is uncertain,  
340 so it is difficult to formulate the real risk when using indoor concentrations of particles. It is  
341 probably important to learn more about the reactive oxygen species and other toxicological  
342 properties of indoor particles. This is particularly true if in a post COVID-world people work from  
343 home and spend more time in domestic settings.

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**Table**

Table 1 Dates of different MCO periods in Malaysia

MCO phase	Dates	Summary
<i>MCO 1</i>	18 <sup>th</sup> March 2020 – 31 <sup>st</sup> March 2020	The prohibition of movement within 10 km of the residence and mass assembly
<i>MCO 2</i>	1 <sup>st</sup> April 2020 – 14 <sup>th</sup> April 2020	The Malaysian government announced a stricter MCO and minimize numbers of essential services
<i>MCO 3</i>	15 <sup>th</sup> April 2020 – 28 <sup>th</sup> April 2020	More roadblocks by police and soldiers
<i>MCO 4</i>	29 <sup>th</sup> April 2020 – 3 <sup>rd</sup> May 2020	Only selected industries are allowed to run full capacity and some movement restrictions were slowly eased
<i>MCO 5</i>	4 <sup>th</sup> May 2020 – 12 <sup>th</sup> May 2020	This phase is known as Conditional Movement Control Order (CMCO). The Government started to ease restrictions on Movement (i.e travelling more than 10 km from residential area is allowed)

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520 **Figure Captions**

521 **Fig. 1.** Map of the study area with monitoring sites marked +.

522 **Fig. 2.** (a) Measured PM<sub>2.5</sub> concentrations as daily averages for the ambient monitoring sites at the  
523 university campus (UPM) and Putrajaya from 9 February to 7 May 2020 (during MCOs phases –  
524 as noted in the Table 1). The open circles denote daytime measurements inside the apartment  
525 (averaged 09:00-19:00) and outdoors as dots. (b) The daily traffic flow at the exit near UPM.

526 **Fig. 3.** Average diurnal change (as an 11-point running mean) in the apartment (dark line) and  
527 outdoors (dotted line) averaged over 13 daytime periods. The grey area denotes the standard  
528 deviation.

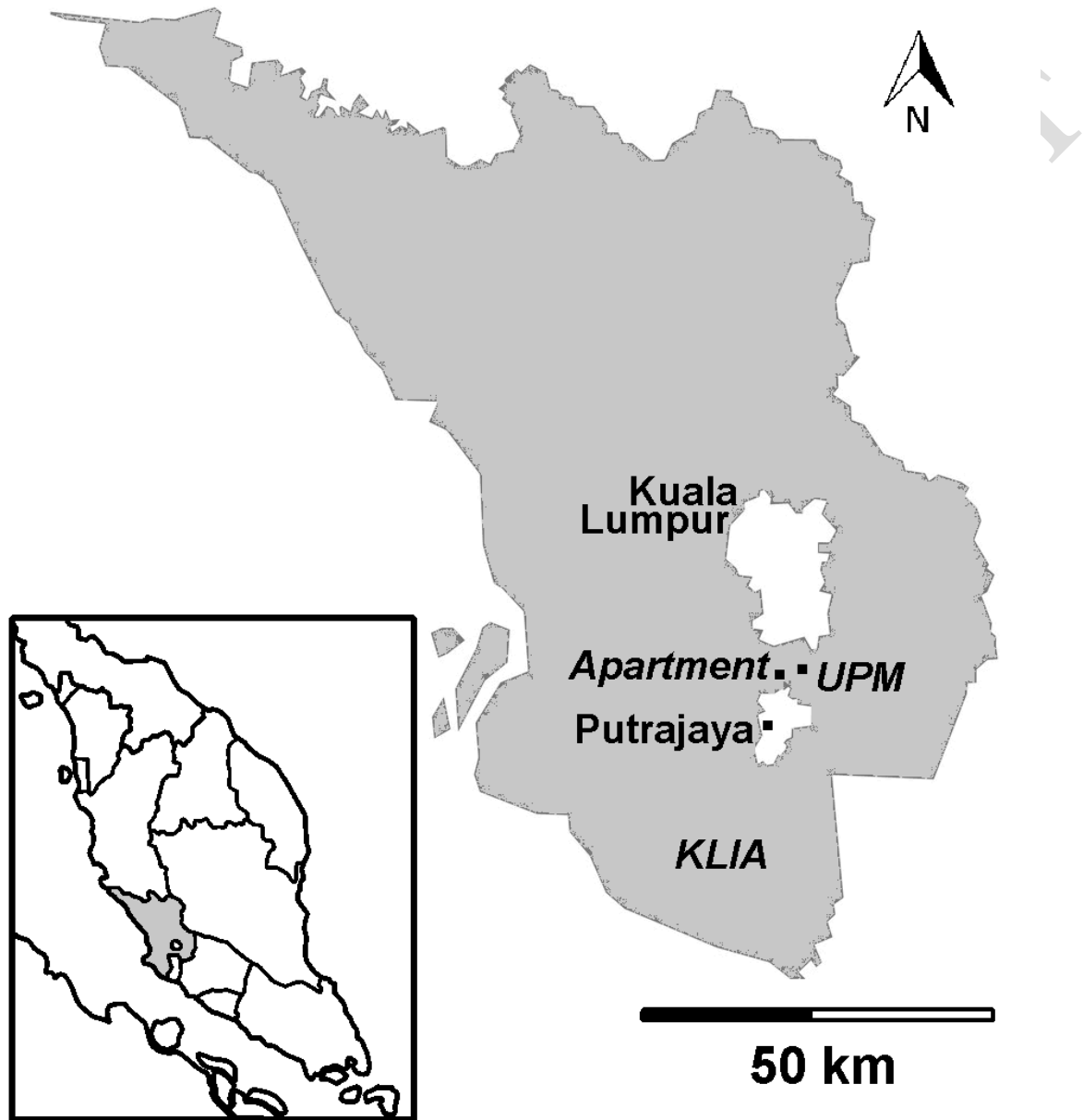
529 **Fig. 4.** The average excess risk of the health burden from acute respiratory diseases (age < 5),  
530 cardio-respiratory diseases (age > 30) and lung cancer (age > 30) estimated for campus and Putrajaya  
531 monitoring station before and during lockdown MCO. The excess risk inside the apartment  
532 assuming concentrations were maintained for 24 hours and 50% assuming they were almost zero  
533 at night.

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Fig. 1.



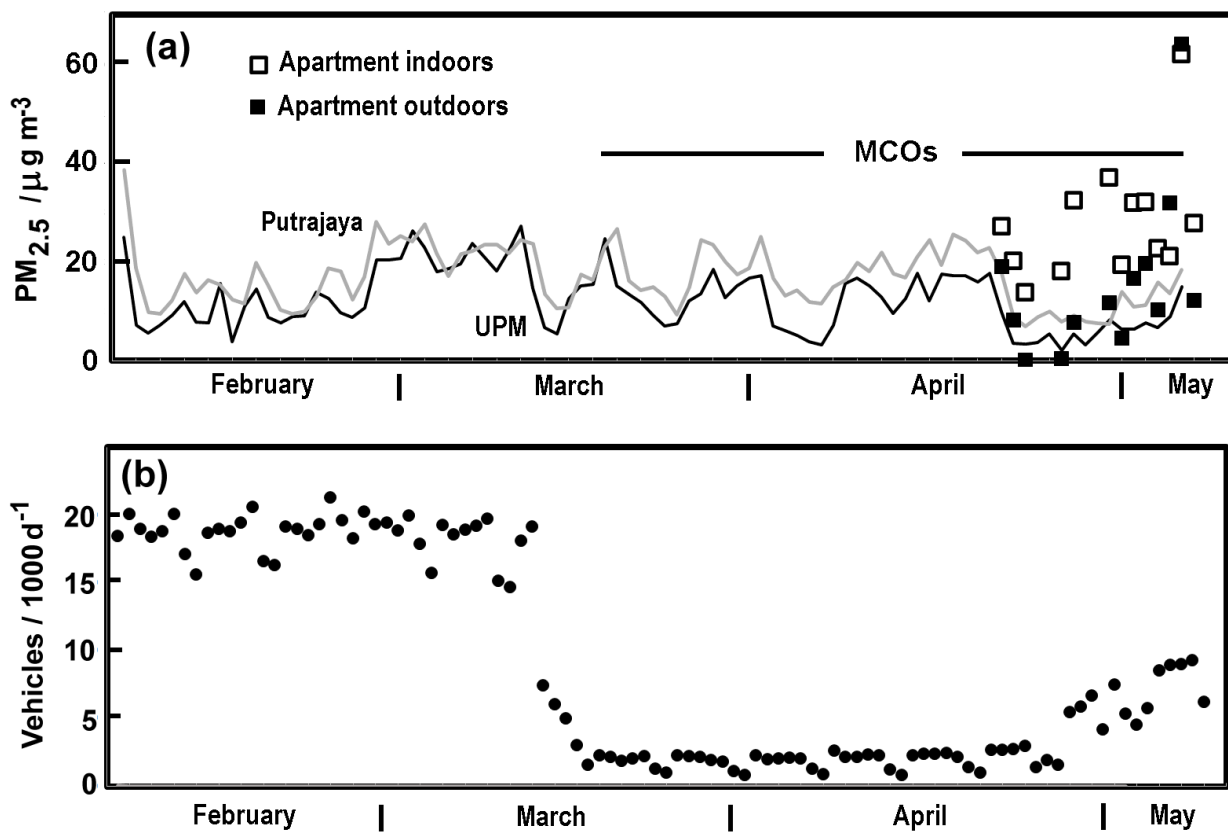


Fig. 2.

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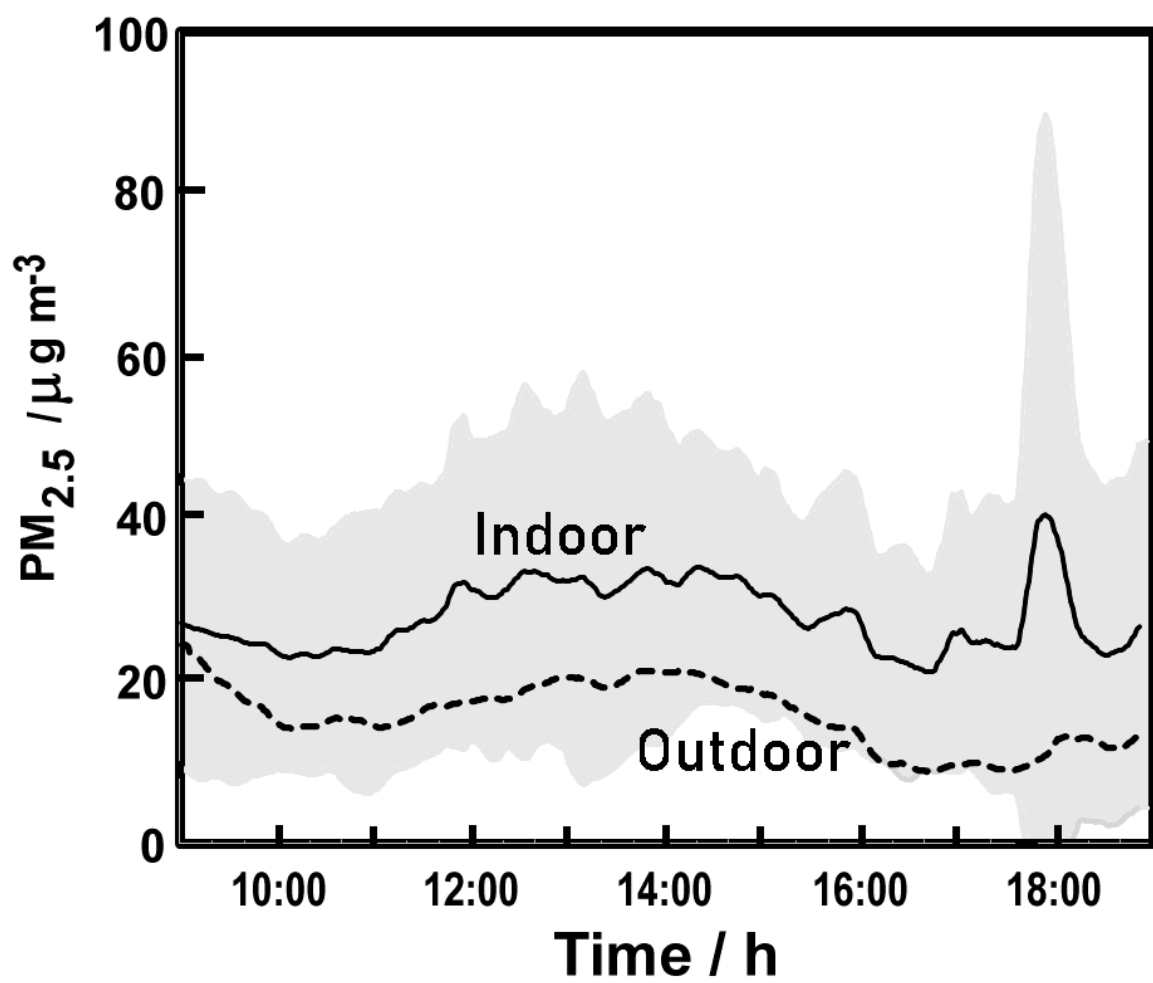
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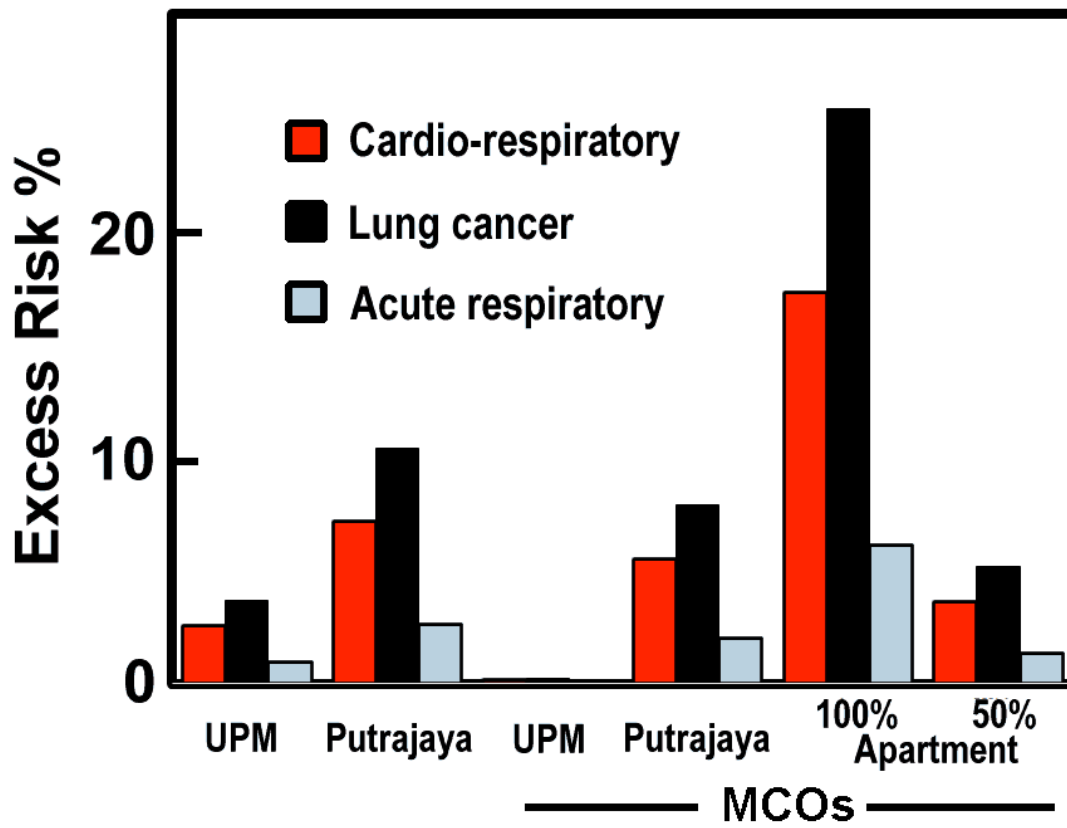
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Fig. 3.



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Fig. 4.